

1 **Modern pollen-vegetation relationships in traditionally mown and**
2 **unmanaged boreal rich-fen communities in central Norway**

3
4 **Kristine Fjordheim**^a, **Asbjørn Moen**^b, **Kari L. Hjelle**^c, **Anne E. Bjune**^{d,e}, **Hilary H.**
5 **Birks**^d

6
7 ^a Department of Biology, University of Bergen, PO Box 7803, N-5020 Bergen, Norway,
8 Kristine.Fjordheim@danielsen-skoler.no

9 ^b Department of Natural History, NTNU University Museum, Norwegian University of
10 Science and Technology, 7491 Trondheim, Norway, asbjorn.moen@ntnu.no

11 ^c Department of Natural History, University Museum of Bergen, University of Bergen, PO
12 Box 7800, N-5020 Bergen, Norway, Kari.Hjelle@uib.no

13 ^d Department of Biology and Bjerknes Centre for Climate Research, University of Bergen, PO
14 Box 7803, N-5020 Bergen, Norway, Anne.Bjune@uib.no, Hilary.Birks@uib.no

15 ^e Uni Research Climate, Nygårdsgaten 112, N- 5008 Bergen, Norway

16

17 **Keywords:**

18 Plant-pollen relationship, pollen representation, pollen surface samples, rich fen, boreal
19 vegetation zone Norway, haymaking

20 **Abstract**

21 Knowledge of the relationship between vegetation and modern pollen deposition is necessary
22 to improve the interpretation of fossil pollen samples. We compared the cover of plant species
23 and the modern pollen deposition in surface moss samples in 49 plots in rich-fen vegetation
24 from the boreal vegetation zone in Norway in two areas (one oceanic and one more

25 continental). These rich fens were used for traditional hay cutting until ca. 1950. Three
26 management regimes are used: 1) biennial mowing 2) quadrennial mowing and 3) unmown.
27 The unmown areas were traditional hay fens, with no management since the 1950s. Of the 89
28 plant taxa and corresponding 65 taxa of pollen and spores, 29 and 22, respectively, were
29 sufficiently represented in both vegetation and as pollen, to allow direct comparisons of the
30 two data sets. Most of the plant and pollen taxa were found in the plots from both study areas
31 and in plots with different mowing frequency. Therefore, quantitative differences in
32 vegetation cover and pollen percentages were the main focus in this investigation of plant-
33 pollen relationships in mown and unmown vegetation. The relative difference in plant cover
34 was largest between biennially and unmown plots. 12 species showed a significant positive
35 correlation with difference in plant cover between mown and unmown areas ($p < 0.05$;
36 Wilcoxon signed rank test), most pronounced for *Molinia caerulea* (most common in
37 unmown plots), and *Carex dioica* and *Thalictrum alpinum* (most common in mown plots). A
38 significant positive plant-pollen correlation (Spearman rank-correlation) was found for 10
39 pollen taxa. *Thalictrum alpinum* is a very good pollen indicator, and Cyperaceae,
40 *Eriophorum*-type and *Pedicularis*-type are good pollen indicators of mowing.
41 Poaceae/*Molinia caerulea* and *Succisa pratensis* are negatively correlated with mowing. The
42 impact of mowing on species composition was similar in both study areas. Mown and
43 unmown areas were differentiated in the vegetation data and could be identified by modern
44 pollen data. These results will improve the interpretation of past land-use practices using
45 pollen analysis of rich-fen vegetation.

46 **1. Introduction**

47 All over northern Europe fens have, for centuries, been important for production of hay for
48 winter fodder. The fertile infields were often of limited extent, so the outfields were of great
49 importance for haymaking, grazing and summer farming (e.g. Ellenberg, 1988; Hjelle et al.,

50 2012; Solem et al., 2012). This long-term regular harvesting has influenced the fens, turning
51 large areas into open semi-natural landscapes. The traditional use of the fens ceased many
52 decades ago in most countries in western and central Europe, including southern
53 Fennoscandia. In parts of central boreal Norway the traditional use of fens, including our
54 study areas, lasted until the 1950s (Moen, 1990; Tretvik and Krogstad, 1999).

55

56 Cultural landscapes leave contemporaneous traces in the form of pollen assemblages in
57 accumulating deposits. Pollen analysis is thus a widely used approach for reconstructing the
58 history of past cultural practices and landscapes (e.g. Berglund, 1991). Investigations of the
59 modern local plant-pollen relationship in areas with documented land-use regimes are a basis
60 for reconstructing these practices back in time. The taphonomy of pollen in modern moss
61 samples is comparable to pollen analytical data from peat cores. Pollen assemblages from
62 vegetation dominated by dwarf-shrubs, graminoids or herbs reflect local sources of non-tree
63 pollen (NAP) and thus small-scale vegetation variations (Hjelle, 1999a; Bunting 2003;
64 Bunting and Hjelle, 2010; Pardoe, 1996). The relationship between number of deposited
65 pollen grains and the corresponding plant cover is complex (e.g. Sugita, 1994), and the pollen
66 taphonomy must be taken into account when interpreting pollen assemblages (e.g. Fægri and
67 Iversen, 1989).

68

69 Mowing leads to disturbance and changes in available resources and competition between
70 species (Crawley, 1997; Grime, 2001). Species resilient to mowing will be favoured and
71 plants with low and rosette growth forms, with meristems close to the ground or clonal
72 growth, are tolerant to mowing (Klimešová et al., 2008). In addition, stress-tolerant species
73 are less affected by mowing (Grime, 2001; Øien & Moen, 2001). Thus mowing affects the
74 composition of vegetation and species abundance. For example the plant cover of *Thalictrum*

75 *alpinum* and several *Carex* and *Eriophorum* species increases with mowing in boreal fens
76 (Moen, 1995).

77

78 In order to reconstruct past vegetation communities from pollen data the relationship between
79 the vegetation community and the pollen produced and deposited within it must be known. If
80 good pollen–plant relationships exist and the plant community is reflecting a type of land use
81 today, pollen types/species may be identified as indicator taxa for that land-use (cf. Behre
82 1981). The present study aims to produce indicator taxa for mown fens which may be of
83 international value when interpreting pollen diagrams. Also the whole pollen assemblage may
84 be used in a comparative approach (Birks and Birks, 1980) to aid in the identification of past
85 vegetation communities or land-use practices. The results from the present study will be
86 combined with plant-pollen data from mown and grazed vegetation types in Western Norway
87 (Hjelle, 1999a) to form a larger data set. This data will aid the interpretation of possible land-
88 use practices (e.g. Gaillard et al., 1994; Hjelle, 1999b) for pollen diagrams from the rich fens
89 in Tågdalen and Sølendet. The data set should also be useful as part of future pollen–plant
90 databases, e.g. the European Pollen Database
91 (<http://www.europeanpollendatabase.net/index.php>).

92

93 The aim of the present study is to assess the modern pollen–plant relationships in rich fens in
94 two study areas. The areas are situated at the transition between the middle and northern
95 boreal vegetation zones, but in different vegetation sections (regional differences oceanity–
96 continentality, Moen 1999). Both study areas were used for traditional haymaking until ca.
97 1950. Regular mowing (with a scythe) of permanent plots was reintroduced in early 1970s
98 and they have been mown regularly for nearly 40 years. The plant communities vary between
99 the study areas, but with the same dominant species, where the impact of mowing on different

100 species is similar (Moen et al., 2012). Further, differences between mowing regimes and
101 unmown areas facilitate various species, giving mainly quantitative differences in plant cover
102 (Moen et al., 2012). We study the correlation between the plants growing in fen vegetation
103 and pollen deposited on a local scale. Three hypotheses are thus put forward: 1) There is a
104 close connection between the taxa in fen vegetation and local pollen deposition, 2) There are
105 small differences in local pollen–plant relationships between the two studied rich fens
106 (regional differences), and 3) Differences between mown and unmown vegetation can be
107 detected from the contemporary pollen assemblages. The modern plant–pollen relationship in
108 fens will provide a basis for interpretations of palaeoecological investigations of past land-use
109 in the study areas.

110 **2. Study areas and plant communities**

111 The two study areas (Fig. 1) are both situated at the transition between the middle boreal and
112 northern boreal vegetation zones of central Norway (Moen, 1999). Tågdalen is an oceanic
113 inner-fjord area nature reserve and Sølendet is a continental nature reserve. Tågdalen is part of
114 the markedly oceanic vegetation section while Sølendet is situated on the transition between
115 the indifferent and slightly continental vegetation sections (sensu Moen, 1999). The 20
116 localities with 49 study plots presented in this paper were established as permanent plots 40
117 years ago, and vegetation and plant populations dynamics are described in a number of papers
118 (e.g. Moen, 1990, 1995, 2000; Aune et al., 1996; Moen et al., 1999, 2012, 2015; Øien and
119 Moen, 2001; Sletvold et al., 2010; Lyngstad et al., 2016). 10 localities are situated along a line
120 over a distance of 940 m in Tågdalen (Fig. 2a). In Sølendet the 10 localities are situated in an
121 area of just below 1 km² (Fig. 2b).

122
123 The duration of the growing season is similar in both areas, from the end of May until the first
124 part of September. Tågdalen has an oceanic climate with high annual precipitation and a

125 thick, long-lasting snow cover, while Sølendet has a more continental climate with less
126 precipitation and cold winters (Table 1). The distance between the study areas is 145 km. In
127 both areas calcareous Cambro-Silurian bedrock (Sigmond et al., 1984) is overlaid with base-
128 rich till (Follestad, 1995).

129

130 The dominant rich-fen vegetation at Tågdalen and Sølendet forms a mosaic with birch
131 woodland (*Betula pubescens*). Sloping fens (slope $>3^\circ$) cover large areas. At Tågdalen the
132 slopes are between 3 and 12° and at Sølendet 3 to 5° . The depth of the underlying peat layer
133 exceeds 50 cm at Tågdalen, and 20 cm at Sølendet. The study areas have the same historical
134 land-use, where the traditional hay cutting declined during the 1930s, and ended in the 1950s.
135 Experimental scything (hereafter called mowing) of permanent plots started in 1973 and was
136 carried out in August, allowing flowering, seed production and dispersal to take place for
137 important species. The localities were established in homogenous fen areas, and the studied
138 plots have been mown regularly biennially or quadrennially since 1970s, or they have been
139 left unmown for ca. 60 years. The hay crop of the studied communities was estimated to about
140 110 g/m^2 and 140 g/m^2 (dry matter) in plots mown biennially and quadrennially, respectively
141 (Moen, 1990; Moen et al., 2015; Lyngstad et al., 2016). Biennial mowing represents the
142 traditional practice, where the harvest output was maximized in relation to labour invested.
143 Quadrennial mowing is a possible equivalent to the mowing frequency during periods of
144 extensive mowing, and during the period when hay-cutting was declining.

145

146 The terminology of mires follows the Fennoscandia tradition in mire ecology (*sensu* Sjörs,
147 1948), separating units related to the main local vegetation gradients. Rich fens are peat-
148 forming mire sites with characteristic vegetation dominated by brown mosses (e.g.
149 *Campylium stellatum*), and with base-rich water (pH above 6). Lawn communities of

150 extremely rich fen vegetation (Sjörs, 1948; Moen et al., 2012; Jiménez-Alfaro et al., 2013)
151 cover the study localities. Phytosociological analyses of the permanent study plots were
152 included in multivariate analyses of 134 rich fen plots from the two study areas (Moen et al.,
153 2012; the plots in this paper mainly belong to their communities II – IV). In the
154 phytosociological classification system (e.g. Dierssen, 1982; Rybniček, 1985), the studied
155 mires belong to the alliances Caricion davalianae Klinka 1934 at Tågdalen, and at Sølendet
156 Caricion atrofuscae Nordh. 1936. The mean number of species in plots with an area of 12.5
157 m² ranged between 30 and 37 (Table 2 and Supplementary A in Moen et al., 2012). The most
158 common vascular plant species in both areas are *Andromeda polifolia*, *Dactylorhiza* spp.,
159 *Equisetum palustre*, *Euphrasia wettsteinii*, *Parnassia palustris*, *Pinguicula vulgaris*,
160 *Potentilla erecta*, *Selaginella selaginoides*, *Succisa pratensis*, *Thalictrum alpinum*, *Tofieldia*
161 *pusilla*, *Carex dioca*, *C. flava*, *C. hostiana*, *C. lasiocarpa*, *C. panicea*, *C. rostrata*,
162 *Eriophorum angustifolium*, *E. latifolium*, *Molinia caerulea* and *Trichophorum cespitosum*.
163 The bottom layer was dominated by *Campylium stellatum*, with *Aneura pinguis*,
164 *Barbilophozia rutheana*, *Fissidens adianthoides*, *Gymnocolea borealis* and *Scorpidium*
165 *cossonii* occurring in all or a large majority of the plots. At Tågdalen the western/lowland
166 species *Drosera longifolia*, *Narthecium ossifragum* and *Schoenus ferrugineus* were present in
167 some plots. At Sølendet a number of alpine/inland species were present, the most common
168 being *Pedicularis oederi*, *Saxifraga aizoides* and *Kobresia simpliciuscula*.

169 **3. Material and methods**

170 *3.1. Research design and vegetation plots*

171 Ten localities with permanent experimental plots with and without mowing within each study
172 area in similar vegetation units of rich fens were chosen for the present research design (Fig.
173 3). Each locality consisted of two or more permanent plots of 2.5×5 m (total 49 plots). The

174 field work for this study was carried out in 2008. Inside each permanent plot one sample plot
175 of 1×1 m was placed in the centre, thus reducing potential edge effects. The minimum
176 distance from a sample plot to an area with different management regime was 0.75 m. The
177 minimum distance between the border of two sample plots was 1.5 m. Initial studies of
178 different sizes of sample plots (from 0.25 m² to 4 m²) revealed that intermediate-sized plots of
179 1 m² were representative of the plant composition of the stand (e.g. Mueller-Dombois and
180 Ellenberg, 1974).

181

182 Three management regimes have been maintained since the 1970s: 1) biennial mowing, 2)
183 quadrennial mowing and 3) unmown since 1950s (Fig. 3). Management regimes 1) and 3)
184 were present in all investigated localities. Six localities at Tågdalen and three localities at
185 Sølendet included management regime 2). For management regime 1) mowing was done one
186 year prior to investigation of plant cover and sampling; for management regime 2) mowing
187 was done three years prior to investigation. The intensive mowing practice 1) represents the
188 traditional mowing, and 2) reflects extensive mowing.

189

190 Plant nomenclature follows Lid and Lid (2005) for vascular plants and Frisvoll et al. (1995)
191 for mosses (Table 2). The species cover was visually recorded using the following cover
192 classes [percent range of cover – cover class mean]: 1 [1-2 plants – 0,25%], 2 [0-1% – 0,5%],
193 3 [1-2,5% – 1,75%], 4 [2,5-5% – 3,75%], 5 [5-10% – 7,5%], 6 [10-20% – 15%], 7 [20-30% –
194 25%], 8 [30-40% – 35%] and so on up to 14. For each taxon the mean plant cover from each
195 management regime was calculated from the average of cover class means from all plots of
196 similar management regime (Table 3). Species recorded in the vegetation are equivalent to the
197 taxa included in “local terrestrial pollen” (see 3.3).

198 3.2. *Surface pollen sampling*

199 Moss polsters were collected from the sample plots as traps for pollen deposition (e.g. Hicks,
200 1977; Hjelle, 1998). Each moss sample consisted of five merged sub-samples, one from the
201 centre of the plot and four in each direction out from the centre and half way to the sample
202 plot border (Fig. 3). The result, a minimum distance of 1 m between a pollen sub-sample and
203 different land-use, reduces the risk of strong plot-edge effects. If moss was absent, sampling
204 was done as close as possible to these points. The modern analogue to a fossil sample from a
205 peat core is a single moss sample, but collecting several sub-samples reduce the potential
206 effect of outliers and of micro-scale differences in pollen content in moss polsters from
207 neighbouring samples (Pardoe, 1996; Hicks, 2001). Moss samples were collected in late July
208 2008. The moss sample consisted of the whole moss turf from the top down to the soil
209 surface. Thus several years of pollen deposition were probably included in the sample (e.g.
210 Mulder and Janssen, 1999; Pardoe et al., 2010), integrating observed year to year fluctuations
211 in fertility and pollen production (Hicks, 2001; Nielsen et al., 2010). Such samples are
212 comparable to the vegetation community when single-year effects can be evened out, as well
213 as to a peat pollen sample reflecting several years. By far the most frequently sampled moss
214 in all localities was *Campylium stellatum*. Among other common species were *Scorpidium*
215 *cossonii*, *Scorpidium scorpioides*, *Drepanocladus* spp. and *Sphagnum* spp.

216

217 3.3. *Pollen processing and analysis*

218 Moss samples were thoroughly rinsed through 450 µm sieves to remove the moss. Pollen
219 preparation of the material passing through the sieve followed standard methods, with KOH
220 and acetolysis (Fægri and Iversen, 1989), but without hydrofluoric acid treatment because the
221 minerogenic content was negligible in all samples. The residue was mounted in glycerol for
222 pollen counting. At least 500 terrestrial-plant pollen (range 563-1349) including more than

223 100 local terrestrial pollen (range 164-724) were counted per sample. Taxa were identified to
224 the lowest possible taxonomic level by the use of keys (Moe, 1974; Fægri and Iversen, 1989;
225 Moore et al., 1991; Beug, 2004) and the modern pollen reference collection at the University
226 of Bergen. Pollen data-analysis was made using TILIA (Grimm, 1990). When possible,
227 Cyperaceae pollen was identified to *Dulichium*-type, representing *Carex* spp., and
228 *Eriophorum*-type pollen (sensu Fægri and Iversen, 1989).

229

230 The percentage calculation sum is the sum of local terrestrial pollen ($\sum LP$). The pollen
231 calculation sum for other pollen and microfossils is $\sum P+X$, where X is the microfossil in
232 question. Local terrestrial pollen and spores represent pollen and spores from taxa belonging
233 to the studied rich fen vegetation (e.g. Janssen, 1966). In this sense extra-local taxa are only or
234 mainly recorded as growing on nearby mineral soil, such as Ericaceae, *Calluna vulgaris*,
235 *Juniperus communis*, *Salix* spp. and *Betula* spp., which occur at low abundances or are absent
236 in the fen plots (Moen et al., 2012). Pollen from *Betula* spp. is mainly a part of the regional
237 pollen deposition, along with pollen from species not represented in the fen localities, like
238 *Pinus sylvestris*. As the yearly moss growth rates may vary between and within moss taxa, as
239 well as with mowing (Moen, 1995), the pollen trapping ability and concentration values will
240 not be directly comparable between samples. Therefore percentages based on the sum of local
241 pollen were used. The use of local terrestrial pollen as the calculation basis will minimize the
242 problems of absolute variations in extra-local and regional pollen related to percentage
243 calculations. Such variations are not related to the vegetation on the fen nor the local mowing.

244 3.4. Data analyses

245 In the data analysis comparing vegetation and pollen data, only the local terrestrial taxa are
246 included. A comparison of species occurrences and pollen data requires comparable
247 taxonomic groups. Therefore, the plant species were grouped according to the pollen taxa that

248 represent them. Analyses of some species were carried out for both single plant species and
249 for the corresponding palynological taxonomic groups, see Table 2.

250

251 *3.4.1. Gradients in the data - ordination*

252 Principal components analyses (PCA) implemented in CANOCO for Windows 4.5 (ter Braak
253 and Šmilauer, 2002) was used to ordinate the 49 vegetation plots according to the pollen taxa
254 they contained. Species centring with square-root transformed data was used. Species scores
255 were divided by their standard deviation to obtain inter-species correlations. Pollen
256 percentages (square-root transformed) from the same plots and taxa and the three
257 management regimes were added as passive samples and environmental variables,
258 respectively, and projected on the ordination diagram. The results were drawn using
259 Canodraw for Windows 4.5 (Fig. 4a and b).

260

261 *3.4.2. Paired comparisons from mown and unmown vegetation*

262 A paired test was carried out comparing the vegetation composition in pairs of mown-
263 unmown plots within each locality. The Wilcoxon signed rank test was used for the
264 nonparametric data to identify taxa with a statistically significant difference in mean
265 percentage plant cover between land-use regimes. It was run in R version 2.11.1 (R
266 Development Core Team, 2010). An exact Wilcoxon signed-rank test was chosen as the
267 grouped data-set has a potential for ties for the pairs. The data from Tågdalen and Sølendet
268 were analysed jointly providing a maximum of 20 pairs. Three sets of tests were run: 1) plots
269 from biennially mown and unmanaged plots, 2) Plots mown biennially and quadrennially, 3)
270 Plots mown quadrennially and unmown plots. Taxa with plant cover registrations in ≥ 6 plots
271 were included.

272

273 *3.4.3. Analyses of differences in species plant cover between management regimes*

274 The indices of difference (ID), equation (1), show whether and how often each taxon
275 increases, displays no change, or decreases in cover from the unmown to the mown pairs of
276 plots:

$$277 \frac{(n_A(1)+n_B(0)+n_C(-1))}{(n_A+n_B+n_C)} \Rightarrow \frac{(n_A-n_C)}{(n_A+n_B+n_C)} \quad (1)$$

278 Where n_A is the number of pairs of plots where each taxon has larger vegetation cover in the
279 mown plot than in the unmown plot, n_B is the number of pairs of plots with no difference, and
280 n_C displays the number of pairs of plots where the taxa has lower vegetation cover in the
281 mown plot than in the unmown plot.

282

283 *3.4.4. Correlation between vegetation and pollen data*

284 To investigate whether plant cover is correlated with the corresponding pollen percentages in
285 the surface samples, the nonparametric Spearman rank-correlation coefficient, ρ (rho) (e.g.
286 Webb et al., 1978), was used to measure statistical dependence between pollen percentages
287 and plant cover for taxa in all plots for the three management regimes. The test was carried
288 out in R version 2.11.1 (R Development Core Team, 2010), examining the absolute
289 percentage differences of paired plots. The statistical significance level was set to 0.05. The
290 data-set of 49 plots from both Tågdalen and Sølendet was used. The correlation test was run
291 for only one study area when plant cover or pollen was absent from the other area. The tests
292 were run for all taxa recorded in ≥ 6 pairs of plots in each paired comparison.

293

294 *3.4.5. Indices of pollen association and representation*

295 Based on presence/absence data of vegetation and pollen from all investigated plots,
296 calculations of indices of association with coincidence (A), over-representation (O) and

297 under-representation (U) (Davis, 1984; Hjelle, 1997; Bunting, 2003; Mazier et al., 2006) were
298 made for each taxon. The definitions of the three indices are:

$$299 \quad A = B_0 (P_0 + P_1 + B_0)^{-1}; \quad O = P_0 (P_0 + B_0)^{-1}; \quad U = P_1 (P_1 + B_0)^{-1} \quad (2)$$

300 where B_0 is the number of plots where the pollen or spore type and the associated taxon is
301 present, P_0 represents the number of plots where the pollen or spore type is recorded but this
302 taxon is not present in the vegetation, and P_1 is the number of plots where the pollen or spore
303 type is absent in the surface sample but the taxon is present in the vegetation. Based on the
304 indices, taxa were grouped into associations: strongly associated types (SAT, $A > 0.65$),
305 associated types (AT, A between 0.65 and 0.5), weakly associated types, (WAT, $A < 0.5$,
306 positive O and U), over-represented type (ORT, $A < 0.5$, $U = 0$), under-represented type (URT,
307 $A < 0.5$, $O = 0$) and unassociated type (UT, $A = 0$, positive O and U).

308 **4. Results**

309 89 plant species (77 vascular plant species and 12 moss species) were recorded from Tågdalen
310 and Sølendet. Taxa present in both the vegetation data-set and the pollen and spore analyses
311 were considered (Table 2). Thus 57 vascular plant taxa were represented in the pollen
312 assemblages and grouped into 26 pollen taxa (Table 2).

313

314 Taxa only recorded in the vegetation, like *Dactylorhiza* spp., *Equisetum* spp. and most moss
315 species, and taxa only identified in the pollen record, e.g. several tree species, were excluded.

316 *4.1. Vegetation data*

317 The most frequent species in both study areas were: *Trichophorum cespitosum* ssp.
318 *cespitosum*, *Thalictrum alpinum*, *Molinia caerulea*, *Potentilla erecta*, *Eriophorum latifolium*,
319 *Carex flava*, *C. panicea*, *C. rostrata* and *Succissa pratensis* (Table 3). *Molinia caerulea*,

320 *Thalictrum alpinum*, *Eriophorum latifolium*, *Betula nana* and several *Carex* taxa displayed a
321 large difference in plant cover between biennially mown and unmown plots (Table 3).

322

323 Twelve plant taxa had significantly different cover values in biennially mown and unmown
324 plots (Table 4 a). Highest cover values in biennially mown plots were found for: four *Carex*
325 species, *Thalictrum alpinum*, *Pinguicula vulgaris*, *Eriophorum latifolium*, *E. angustifolium*
326 and *Pedicularis palustris*. *Molinia caerulea* was much more abundant in unmown plots, and
327 this was the only plant taxon displaying a difference in plant cover between all the
328 management regimes. *Betula nana* and *Succisa pratensis* were more abundant in unmown
329 plots than in biennially mown plots. Combining the plant taxa into pollen groups resulted in
330 *Dulichium*-type (i.e. *Carex* spp.) displaying a difference in vegetation cover between all
331 management regimes, whereas *Eriophorum*, Cyperaceae and Poaceae only displayed a
332 difference between the two mown and the unmown plots (Table 4 b). The difference in plant
333 cover in biennially mown and unmown plots represented the main variation in plant cover in
334 fen vegetation (Table 4 a), most pronounced for *Molinia caerulea*, *Carex dioica* and
335 *Thalictrum alpinum*.

336 4.2. Ordination of vegetation and pollen data

337 The first two axes of the PCA-ordination of the 49 plots (Fig. 4) had eigenvalues of 0.25 and
338 0.23 respectively. The other axes had eigenvalues lower than 0.10 (not shown). The
339 vegetation samples representing different mowing frequencies were spread along the whole
340 range of the first axis (Fig. 4a), but were separated on the second axis, reflected in both the
341 dispersal of samples and the mean scores for the environmental variables. On axis 2 the
342 variable representing biennial mowing had a high negative mean score, while quadrennial
343 mowing had a small negative mean score and unmown had a high positive mean score. The
344 majority of pollen samples (passive data) had low scores on both axes. The pollen samples

345 showed a similar pattern as the vegetation samples, where the pollen samples from the
346 unmown plots to a large extent are separated from the plots with the two mowing frequencies,
347 most pronounced for unmown and biennially mown plots.

348

349 *Pedicularis*-type had a negative score on the first axis of the PCA ordination of plant taxa (as
350 pollen taxa groups, Fig. 4b) whereas *Potentilla erecta*, *Rhinanthus*-type and *Succisa pratensis*
351 had high scores. *Dulichium*-type, *Thalictrum*, Cyperaceae and *Eriophorum* had lowest scores
352 on the second axis whereas Poaceae had the highest score followed by *Sphagnum*.

353 4.3. Pollen data in relation to management regimes

354 Of the 65 palynological taxa found at Tågdalen and Sølendet, only 26 pollen and spore taxa
355 represented the investigated, contemporary vegetation (Table 2). Pollen taxa present in >4
356 plots are displayed in Table 5.

357

358 At Tågdalen *Thalictrum* and *Eriophorum*-type pollen had their largest pollen percentages in
359 mown plots, especially seen for *Thalictrum* in biennially mown plots in Tågdalen. Poaceae
360 and *Potentilla*-type had largest pollen percentages in unmown plots. *Dulichium*-type pollen
361 had largest pollen percentages in unmown plots, but in Tågdalen the differences are small.
362 Cyperaceae had generally small relative changes to no differences between management
363 regimes, and the standard error is large. *Succisa pratensis* had also small to no differences
364 between management regimes in Tågdalen, where at Sølendet the largest pollen percentages
365 were registered in unmown plots. *Pedicularis*-type pollen had a general low pollen presence.
366 The *Pedicularis*-type pollen percentages were slightly higher in biennially mown than in
367 unmown plots in Sølendet, and where absent in unmown plots in Tågdalen.

368

369 From Sølendet only three plots represent quadrennially mown fen, giving a potential for
370 uncertainties regarding the result. In Sølendet the pollen data does not always follow the trend
371 from Tågdalen, as for *Thalictrum* and Poaceae.

372

373 4.4. Pollen-vegetation relationships

374 Spearman rank-correlation coefficients relating plant cover to pollen percentages are
375 presented in Table 6, along with the indices of association and association types which reflect
376 relationships between presence/absence of plants and their pollen. As *Molinia caerulea*
377 formed the majority of the Poaceae plant cover (Table 3), both *M. caerulea* and Poaceae plant
378 cover were individually correlated to the Poaceae pollen data. *Eriophorum*-type pollen was
379 compared to both *Eriophorum* spp. and *Eriophorum*-type vegetation.

380

381 Most of the strongly associated plant-pollen types (SAT) and the associated plant-pollen type
382 (AT) had a statistically positive correlation between plant cover and pollen percentages (Table
383 6). These were Rosaceae, *Potentilla*-type, *Pedicularis*, *Thalictrum/Thalictrum alpinum*,
384 *Selaginella selaginoides*, *Eriophorum*-type/*Eriophorum* spp., Poaceae/*Molinia caerulea* and
385 Cyperaceae. *Succisa pratensis* and *Menyanthes trifoliata* had significant plant-pollen
386 correlations, but had weak associations with their corresponding pollen types as they
387 displayed a trend towards pollen under-representation.

388

389 The remaining taxa in Table 6 did not have a significant relationship between the vegetation
390 cover and pollen percentages. However, common species in the vegetation, such as Poaceae,
391 *Eriophorum* spp. and *Carex* spp., were also common in the pollen assemblages, reflected in a
392 strong association. Underrepresented taxa were insect pollinated species which were poorly
393 presented in the pollen assemblages, but they were present in the vegetation when present as

394 pollen, e.g. *Drosera rotundifolia*-type, *Bistorta vivipara*-type and *Saxifraga oppositifolia*-
395 type.

396 **5. Discussion**

397 Rich fen vegetation is widely distributed in Europe. Vegetation plots from our two study areas
398 are included in a European study, and are classified as boreal fens (Jiménez-Alfaro et al.,
399 2013, cluster 3a). The relationships between vegetation and pollen assemblages in our two fen
400 areas add to the database of surface-pollen samples from a variety of vegetation types and
401 land-use regimes (e.g. Gaillard et al., 1994; Hjelle, 1999a; Mazier et al., 2006; Ejarque et al.,
402 2011; Waller et al., 2017). These data contribute to modern-pollen studies and long-term
403 investigations. However, there are few studies that consider rich fens. An exception is the
404 nemoral rich fens of Waller et al. (2005, 2017) which include a number of species in common
405 with our studied fens; e.g. *Potentilla erecta*, *Carex nigra*, *C. panicea*, *C. rostrata* and *Molinia*
406 *caerulea*. These are among the most common species in our two study areas as well as in the
407 two study areas in southern and eastern England. The annually cut “Sedge fen” included in
408 Waller et al. (2017) has many features in common with our fens, and represent a community
409 with close relationship between vegetation and local pollen deposition.

410

411 Comparable to our study, Waller et al. (2017) found that nemoral rich fen vegetation subject
412 to different types of management, including different cutting regimes, produce distinctive
413 pollen signatures. Their study area is in nemoral vegetation zone, with somewhat different
414 plant communities and species, compared to our boreal rich sloping fens. Also, de Klerk et al.
415 (2017), in arctic fens, state that pollen deposition reflects vegetation, and that regional pollen
416 deposition relate to different biogeographical regions.

417 *5.1. Plant cover and pollen deposition*

418 The pollen recovery from moss samples is a measure of plant representation based on fertility
419 and pollen production, and not only plant cover. However, there is a close connection
420 between plant cover, plant biomass and flowering of a large number of species, as shown in
421 the long-term population and vegetation studies in the two study areas (Moen, 1990, 1995;
422 Aune et al., 1996; Lyngstad et al., 2016). Mowing leads to a reduction in above ground
423 biomass and plant cover and a reduction in flowering for several tall-growing species, e.g.
424 *Molinia caerulea* and shrubs (Table 3). A number of low growing species, with most of the
425 above-ground biomass close to the surface (not reached by the scythe) increased in biomass,
426 cover and flowering, e.g. *Thalictrum alpinum* and *Carex dioica*. Thus pollen recovery is
427 regarded an approximation of plant cover of a taxon.

428

429 There can be a large variation in annual flowering, pollen production and pollen deposition
430 between years (e.g. Hicks, 2001; Hättestrand et al., 2008), and several years of pollen
431 deposition should be sampled (Pardoe et al., 2010). Although moss samples may contain only
432 one or two years of pollen deposition (Räsänen et al., 2004), it is generally assumed that moss
433 polsters contain several years of pollen deposition (Bunting et al., 2013). The annual pollen
434 productivity for species of boreal trees such as *Pinus*, *Betula*, and *Picea* has been found to be
435 related to summer temperature of the year prior to pollen emission (Autio and Hicks, 2004).
436 Our moss samples were gathered in late July 2008. The meteorological stations in the two
437 study areas recorded summer temperatures close to the 30-year normal for the summers 2007
438 and 2008 (Lyngstad et al., 2016). Both 2007 and 2008 were normal/good flowering years for
439 most of the studied species; e.g. for *Eriophorum latifolium* (Lyngstad et al., 2016), indicating
440 that even if only a few years are represented in the moss samples, the pollen data are probably
441 representative.

442

443 The investigated plots are designed for vegetation studies (e.g. Lyngstad et al., 2016), where
444 different management regimes (and collected moss polsters) are very close, but still there
445 were differences between pollen assemblages taken 1-4 m apart. The differences in pollen
446 percentages corresponded well with the vegetation from different management regimes,
447 indicating that a substantial part of pollen was derived locally and that rich fens are suitable
448 for local-scale plant-pollen studies. Comparable studies have also demonstrated a strong
449 correspondence between local vegetation and pollen assemblages for herbs and dwarf-shrubs
450 (e.g. Pardoe, 1996; Hjelle, 1999a; Waller et al., 2017) and a relevant pollen source area (*sensu*
451 Sugita, 1994) of only a few meters is found within mire communities (Bunting 2003; Bunting
452 and Hjelle, 2010).

453 5.2. Identification of management regimes by modern pollen assemblages

454 Most plant species occurred in both study areas, and in plots with different mowing regimes.
455 Those with high cover in all plots were *Trichophorum cespitosum*, *Thalictrum alpinum* and
456 *Molinia caerulea*. Six other vascular plant species occurred in at least 80% of the plots: *Carex*
457 *dioica*, *C. panicea*, *Eriophorum angustifolium*, *E. latifolium*, *Potentilla erecta* and *Selaginella*
458 *selaginoides*. Many bryophytes occurred in the plots, e.g. the dominant and constant rich-fen
459 species *Campylium stellatum* and *Scorpidium cossonii* (Moen et al., 2012). Several rare
460 species distinguish between the fens of the two areas, leading to the classification of the
461 oceanic and the more continental fens in different alliances in the phytosociological system
462 (Moen et al., 2012). All of these species had low pollen percentages, and they are not included
463 in this discussion.

464

465 The differences in plant cover and pollen assemblages between the management regimes were
466 mainly quantitative. Earlier studies from these localities have shown that the variation in

467 vegetation cover of plots from similar land-use between localities can be larger than between
468 plots from different management regimes within the same locality, and that permanent plots
469 with different management regimes inside each locality belong to the same plant
470 community/vegetation cluster (Moen et al., 2012; Lyngstad et al., 2016). The differences
471 between vegetation plots from different management regimes were more pronounced than
472 the differences between pollen samples from these regimes (Fig. 4a). Generally, the plant
473 compositions in plots mown biennially and quadrennially were more similar than to unmown
474 vegetation (Fig. 4a and b), whereas this to a lesser degree was found for pollen data (Fig. 4a).
475 Similar results for plant cover were found in the same areas (e.g. Aune et al., 1996), as well as
476 for plant and pollen data from other studies of mown vegetation (e.g. Hjelle, 1999a).

477

478 Palynological studies of rich fens have received little attention in comparison to lakes and
479 raised bogs (Waller et al., 2017). Our results are a contribution to knowledge established in
480 earlier studies from fen vegetation (e.g. de Klerk et al., 2017; Waller et al., 2017)
481 demonstrating that fen vegetation produce distinctive pollen signatures with indicator taxa
482 and should be used also for palaeoecological studies.

483

484 To detect past types of land use and changes in land use by pollen analysis, taxa that show
485 responses to land use regimes, in the present case mowing, and have pollen deposition that
486 correlate with plant cover, must be identified. Here five groups of pollen taxa with different
487 potential for land-use reconstruction were identified (Table 7).

488

489 *5.2.1. Very good and good pollen indicators for management regimes*

490 In rich fens *Thalictrum* is sensitive to land-use changes, and the contemporary pollen
491 percentages correlate with the local plant cover. *Thalictrum alpinum* occurred in all plots,

492 with highest cover in biennially mown plots compared with unmown plots (Fig. 4b),
493 absolutely (Table 3) and relatively (Table 4a). This is in accordance with previous studies
494 from Sølendet, which also report increased biomass, cover and flowering with mowing
495 frequency (Moen, 1995; Aune et al., 1996). A difference in cover of *T. alpinum* between
496 biennially and quadrennially mown plots is not clear from this study. Its growing point is at
497 the soil surface and is little affected by mowing, so it becomes more abundant in mown fens
498 where competition is reduced. The effect of competition was demonstrated by Klanderud and
499 Totland (2005), who removed neighbouring vegetation in an alpine community and found
500 increases in the number and cover of leaves and reduction in the length of the flowering stems
501 of *T. alpinum*. However, *T. alpinum* is also present in grazed communities in outfields, as in
502 Budalen, located between our two study areas (Austrheim et al. 1999), and without comparing
503 mown and grazed communities one cannot exclude *Thalictrum* as indicator of more than one
504 land-use practice. The strong association of the presence of both plant and pollen and absence
505 of pollen when plants are lacking, indicates a limited pollen-dispersal distance from these
506 plots with a minimum distance of 1 m. *T. alpinum* is wind dispersed, but dispersal can be
507 limited by the short flower stems. This means that *Thalictrum* pollen is a good indicator of
508 local plant presence, and an increase in plant cover can indicate land use like mowing (Table
509 7). Increased percentages of *Thalictrum* pollen in peat profiles have been used as an indicator
510 of mowing in vegetation history studies of rich fens in boreal areas of central Norway
511 (Gunnarsdóttir, 1999; Solem et al., 2012).

512

513 The interpretation of the pollen representation of Cyperaceae undiff. is more challenging. The
514 family is generally well adapted to mowing, although with variations and opposite trends for
515 single species, as for *Carex* species (Table 4 a) (e.g. Moen, 1990, 1995). Both mowing
516 regimes differed from unmown plots in total Cyperaceae plant-cover percentages, but the

517 differences between mowing intensities could not be detected in modern pollen samples. The
518 differences in the Cyperaceae pollen percentages between management regimes are small.
519 The strong plant-pollen correspondence from these closely situated plots suggests a limited
520 dispersal distance for Cyperaceae, which was reported by Bunting and Hjelle (2010) to be
521 between 1.5 and 3.5 m. The present investigation has shown the value of separating
522 *Dulichium*-type pollen, which in our area consists of *Carex* spp., from Cyperaceae undiff.
523 pollen.

524

525 *Eriophorum*-type pollen includes three *Eriophorum* species and two *Trichophorum* species.
526 *Eriophorum* spp. and *Trichophorum* spp. (included in Cyperaceae) were most abundant in
527 vegetation mown biennially (Fig. 4b). *E. angustifolium*, *E. vaginatum* and the more
528 widespread species *E. latifolium*, were well adapted to mowing, which also promoted
529 increased flowering according to Moen (1990, 1995), Aune et al. (1996) and Lyngstad et al.
530 (2016). *Trichophorum cespitosum* ssp. *cespitosum*, the dominant species in the rich fens, was
531 indifferent to mowing, as found in earlier studies (e.g. Moen, 1995; Aune et al., 1996) and the
532 difference between management regimes was relatively small.

533

534 There was a positive plant-pollen correlation between *Eriophorum*-type pollen and total
535 *Eriophorum* species, but not for the pollen type and all five individual species. Most of the
536 pollen type might be from *Eriophorum* if *Trichophorum* pollen deposition was reflecting the
537 small variation in *Trichophorum* vegetation cover. The difference in *Eriophorum* spp. plant
538 cover between the two mowing regimes and unmown plots could be detected from pollen
539 analysis, as found from other studies (Hjelle, 1998), indicating that *Eriophorum*-type pollen
540 should be separated from Cyperaceae undiff. if possible to increase its value in the
541 interpretation of past communities.

542

543 *Molinia caerulea* is the most abundant grass in the plots but the pollen taxon Poaceae can also
544 include *Deschampsia cespitosa*, *Festuca ovina* and *Nardus stricta*. *M. caerulea* is clearly
545 most abundant and very fertile in unmown plots (Moen, 1990, 1995; Aune et al., 1996). The
546 plant cover of *M. caerulea* decreased with the intensity of mowing, whereas the cover of
547 Poaceae only differed clearly between the mowing regimes and unmown plots (Table 3 and
548 4). The main reason for the decrease of *M. caerulea* in mown fens is the exposure of the
549 elongation zone to the mower (scythe), strongly reducing the plants. Under moist conditions
550 grazing also reduces plant cover of *M. caerulea* (Hume et al., 1999). In boreal shrub-
551 dominated plant communities and in areas with mowing every 5-10 years (Aune et al., 1996)
552 and in temperate tall-growing fen communities (Rowell et al., 1985) *M. caerulea* increases in
553 biomass and cover with mowing. On the generally nutrient-poor boreal fens *M. caerulea*
554 suffers from the removal of reserves by intensive cutting during a short growing season (Øien
555 and Moen, 2001). In more fertile localities, strong competitors like shrubs and tall herbs are
556 removed by mowing. In such areas, and under mowing with higher stubble, the *Molinia*
557 tussocks can develop. In addition, the fertility of *M. caerulea* is reduced in mown plots on
558 boreal rich fens (Moen, 1995). The modern Poaceae pollen percentages correlated with the
559 plant cover of *Molinia*, but not with all Poaceae species (Table 6). This indicates that most of
560 the Poaceae pollen originated from *M. caerulea*. *Molinia caerulea*/Poaceae displays a
561 difference in plant cover and pollen percentages between areas mown biennially and
562 quadrennially from Tågdalen, whereas the data from quadrennially mown areas in Sølendet
563 are few and difficult to interpret. Overall, mowing of the rich fens results in a reduction of
564 Poaceae plants and pollen and an increase of Cyperaceae plants and pollen. This is in
565 accordance with investigated fen communities in UK where Cyperaceae (including *Carex*
566 *nigra*, *C. panicea* and *C. rostrata*) have high values in intensively managed communities and

567 decrease with longer cutting rotation when Poaceae (including *Molinia*) increases (Waller et
568 al., 2017). The results from fens seem therefore to contrast to pollen assemblages from dry
569 meadows dominated by high Poaceae pollen values and relatively low values of Cyperaceae
570 (Hjelle, 1999a).

571
572 *Succisa pratensis* was present throughout the rich fens, with its largest cover in unmown
573 plots, as reported in Moen (1995) (Fig. 4b, Table 3 and 4). In other communities a decrease in
574 competition following mowing allows an increase in plant cover of *S. pratensis* (Moen 1995).
575 However, with relatively little competition in the boreal rich-fens the effect of reduced
576 biomass had an important impact on this species. In mesic to dry grasslands *S. pratensis*
577 increase in cover with grazing (Herben et al., 2006), making both vegetation type and type of
578 land use important for *S. pratensis* plant cover. The cover of *S. pratensis* in the plots was low
579 and the differences between management regimes were small. It is insect-pollinated and has a
580 limited dispersal distance (Hjelle, 1997, 1998), reflected also in this study by a strong plant-
581 pollen relationship, but with a tendency of being under-represented. This makes the species
582 suited for interpretation of local vegetation, but less important for differentiating the
583 investigated management regimes in the present study.

584
585 *Pedicularis*-type pollen included *Pedicularis palustris* and *P. sylvatica* at Tågdalen and
586 mainly *P. palustris* and *P. oederi*, in addition to the rare *P. sceptrum-carolinum* at Sølendet
587 (Moen, 1990, 2000). The pollen percentages of *Pedicularis*-type related well to the cover for
588 *Pedicularis* spp., possibly because *P. palustris* is the main pollen source as well as the main
589 plant in the studied vegetation. As an associated type *Pedicularis* species were common, but
590 not dominant in either vegetation or in pollen deposits. Both *Pedicularis* spp. and *P. palustris*
591 had greatest cover in mown plots, with a trend to greater abundance in quadrennially mown
592 plots. Only biennially mown and unmown plots displayed a difference even if the percent

593 values are small. Earlier studies (e.g. Moen, 1995) confirm this, along with increased
594 flowering frequency in mown plots. *Pedicularis* has a very low pollen representation, so even
595 small differences in pollen percentages may represent substantial differences in number of
596 plants between plots. *P. palustris* is a biennial species mainly occurring in mud bottoms and
597 carpets with a scattered field layer, i.e. localities with low competition. *P. oederi*, a common
598 species in lawns and open fen margins at Sølendet (Moen, 1990) occurred in only 11 of 49
599 plots. This species shows reduced plant biomass and lower flowering frequency in response to
600 mowing, especially intensive mowing, and is also reduced by abandonment during
601 succession. The pollen data showed a correlation with plant cover, in contrast to the usually
602 rare occurrence of *Pedicularis* in pollen diagrams. Bunting (2003) found no association
603 between plant and pollen in surface studies of heathland communities in UK. Our results
604 indicate that *P. palustris* has a larger potential of being recorded than the other species, which
605 may indicate that *Pedicularis* may be regarded as an indicator of mown fens.

606

607 There are short distances between areas of different land-use in the investigated plots. The
608 strong plant-pollen relationship and strong association of presence indicate a local pollen
609 source area of approximately 1 to a few meters for taxa defined as very good and good pollen
610 indicators for managerial regimes.

611

612 5.2.2. Potential pollen and spore indicators for managerial regimes

613 Within the homogenous rich-fen community a wide distribution of both plants and pollen
614 gave a high probability of association for many taxa (cf. Hjelle, 1997; Mazier et al., 2006).
615 However, for several of these taxa there was no identified plant response to different
616 managerial regimes. From other studies *Potentilla erecta* displays larger cover and
617 increased flowering in unmanaged plots compared to mown fens (Moen, 1995; Aune et al.,

618 1996). A reduction in plant cover of *P. erecta* has been documentet from grazed vegetation on
619 moist soil (Hulme et al., 1999). With the high cover, the relatively coarse resolution for
620 abundant taxa in the chosen cover scale (e.g. Bunting and Hjelle, 2010), and substantial
621 variations between and within managemental regimes, potential vegetation trends were not
622 identified in the present study (Fig. 4b). Rosaceae undifferentiated pollen correlated with *P.*
623 *erecta*, its only species in the vegetation, thus Rosaceae pollen was interpreted to mainly
624 representing *P. erecta*. With a good plant-pollen correlation *P. erecta* reflects local
625 vegetation, it is a good indicator of grazing (e.g. Hjelle, 1999; Mazier et al., 2006) and can, in
626 the present study, represent a pollen indicator type of open vegetation more than mowing.
627 High moisture levels might limit the plant cover of *P. erecta* on these rich fens, playing a
628 more important role in plant distribution than land use.

629

630 *Selaginella selaginoides* is known to be favoured in mown fens (Moen, 1995; Aune et al.,
631 1996). In the studied fens the spore recovery was a good indicator of its plant presence, but no
632 statistical relationship between plant cover and management regime has been identified. With
633 low percentages on the investigated fens, any trends have been difficult to decipher. There
634 was a tendency of increased plant presence with mowing from pairwise comparisons (Table 3
635 and 4), even if the general tendency in the plots is a slight decrease with mowing (Fig. 4b). If
636 its land-use response is identified from further studies, *S. selaginoides* could be an indicator-
637 type of mowing.

638

639 5.2.3. Pollen types reflecting unidentified factors or not reflecting plant cover

640 *Menyanthes trifoliata* was rare but still demonstrated a positive plant-pollen correlation on
641 these rich fens. Its main distribution on the fens was outside the investigated localities, in the
642 wettest places in the fens and in small lakes (Moen et al., 2012). However, in lawn and carpet

643 communities, the cover of *Menyanthes* increases and flowering decreases with mowing
644 (Moen et al. unpublished data).

645

646 Two sub-groups of pollen types not reflecting plant cover were identified:

647 a) Mowing regime affected the species cover, but there was no significant plant-pollen
648 correlation. The strong association between plant and pollen was due to their large abundance.
649 *Carex* species display a variety of plant responses to mowing regarding biomass, cover and
650 fertility (Moen, 1995). Several *Carex* species and the genus as a whole were most abundant in
651 mown plots (Fig. 4b; Moen, 1990; Aune et al., 1996). The low taxonomic precision of
652 *Dulichium*-type pollen and variations in pollen production within the genus (Randall et al.,
653 1986; Hjelle, 1998) were probable reasons for a lack of plant-pollen correlation. *Dulichium*-
654 type pollen has a potential in palaeoreconstruction if the plant-pollen relationship can be
655 interpreted locally. *Betula* species as a whole, and the main species *B. nana* and *B. pubescens*
656 were mainly found in relation to succession on the fen, along the margins and as surrounding
657 vegetation (e.g. Moen, 1990). The over-represented and variable pollen recovery from *Betula*
658 was probably from the extra-local and the regional vegetation. Thus *Betula* pollen has
659 limitations in reconstructing local vegetation, but an increase in pollen may indicate fen
660 succession and increased tree cover on extra-local and regional scales. The cover of *Drosera*
661 spp. at Tågdalen and *Pinguicula vulgaris* at Sølendet increased with mowing, as shown by
662 Moen (1995). As entomophilous and cleistogamous species, respectively, their strongly
663 under-represented and unassociated pollen document the presence of the plant, but they are
664 not suitable to represent variation in the vegetation.

665

666 b) The second sub-group includes taxa with no identified plant response to different land-use
667 and no plant-pollen correlation on the rich fens (Table 7). These were rare in this vegetation

668 type, but could be common in nearby herb and heather woodland (Moen, 1900, 2000). The
669 pollen might mainly be extra-local. These taxa are of limited value in local land-use
670 reconstructions in such boreal fens, but can be valuable in reconstructions of extra-local
671 vegetation, succession and fen dynamics.

672

673 Changes in climate and fen characteristics might change the effect mowing plays on single
674 taxa through time as well as geographically. Still, the ecological effect of mowing will be
675 similar for many taxa under various conditions, and thus comparable between regions (e.g.
676 Hjelle, 1999a).

677 **6. Conclusion**

678 The differences in plant cover between mown and unmanaged plant communities on the
679 boreal rich-fens were mainly quantitative, where the same taxa are present in the vegetation in
680 mown and unmown fens, but in varying amounts (Table 3). For several taxa managemental
681 regimes explained a substantial part of the plant-cover variation on the boreal rich-fens. The
682 relative changes of several taxa in the vegetation rather than the presence or absence of any
683 single pollen-indicator taxon indicate land-use changes on these fens.

684

685 A correlation between the plant cover of taxa constituting the majority of the vegetation on
686 the rich fen and their contemporary, deposited pollen on a local scale was documented. The
687 short distances between different managemental regimes in this study indicate a local pollen
688 source area from 1 and up to a few meters for important taxa in the mown fens, like
689 *Thalictrum alpinum*, Cyperaceae and *Succisa pratensis*. Thus, presences of their pollen in
690 peat samples can generally be assumed to reflect presence in the vegetation.

691

692 Modern pollen assemblages, with correlation to plant cover, display a difference in pollen
693 percentage between biennially mown and unmown vegetation. Further, several taxa have a
694 potential for indicating mowing quadrennially, even if the two mowing regimes are not
695 clearly separated in the present study. A modern pollen assemblage indicating mowing on
696 these rich fens has high percentages of *Thalictrum*, *Eriophorum*-type and Cyperaceae undiff.
697 pollen. *Pedicularis* pollen percentages are generally higher than in unmown vegetation, and
698 Poaceae (*Molinia caerulea*) and *Succisa pratensis* pollen percentages are lower than in
699 unmown vegetation. *Potentilla*-type and Rosaceae, pollen together representing *Potentilla*
700 *erecta*, decrease with mowing of the fens, whereas an increase in *Selaginella selaginoides*
701 spores potentially indicates mowing. In combination, these changes in pollen percentages
702 from such species can be considered a fingerprint for mowing regimes on rich fens. In other
703 vegetation types and under different land use, some of these species would display a different
704 combination of increase and decrease in plant cover and pollen deposition.

705

706 This study emphasises the importance of identification of pollen to as low a taxonomic level
707 as possible. The separation of *Eriophorum*- and *Dulichium*-type (i.e. *Carex* spp.) pollen from
708 Cyperaceae undiff. is essential as the two former, partly in different ways, are important taxa
709 in identifying land-use regimes.

710

711 This study has shown a close connection between taxa in fen vegetation and the local pollen
712 deposition in the two contrasting climate regimes (oceanic and continental) of the middle and
713 northern boreal vegetation zones, confirming hypothesis 1. Boreal rich-fens in the two areas
714 have the same dominant species and the same species producing the majority of the pollen,
715 supporting hypothesis 2. The difference between mown and unmown vegetation can be
716 detected from the contemporary pollen assemblages, particularly between biennial mown and

717 unmown vegetation, supporting hypothesis 3. In this study *Molinia caerulea*/Poaceae display
718 a difference in plant cover and pollen percentages between areas mown biennially and
719 quadrennially. This difference between biennially and quadrennially mown areas was not
720 identified for most taxa, and should be further investigated. The results from the present paper
721 contribute to a better understanding of vegetation-pollen relationships on rich fens and
722 separation of mown and unmown fens, based on pollen assemblages. Our findings will aid in
723 the interpretation of fossil pollen assemblages in terms of local land use on fens in the boreal
724 zone.

725

726 **Acknowledgements**

727 We are grateful to John Birks for valuable comments on the research design and to Anders
728 Lyngstad for help during field work, and Dag-Inge Øien for preparing the maps from Tågdalen
729 and Sølendet. The manuscript was improved by the valuable suggestions and comments from
730 two anonymous referees. This study has received financial support from The Olaf Grolle
731 Olsen and Miranda Bødtker legacy and A. Heiberg and H.B. Fasmers fund, both at the
732 University of Bergen. This study was part of a PhD project at the Department of Biology,
733 University of Bergen.

734 **References**

- 735 Aune, E.I., Kubiček, F., Moen, A., Øien, D-I., 1996. Above- and below-ground biomass of boreal outlying
736 haylands at the Sølendet Nature Reserve, Central Norway. *Norwegian Journal of Agricultural Sciences*
737 10, 125-152.
- 738 Austrheim, G., Olsson, E.G.A., Grøntvedt, E., 1999. Land-use impact on plant communities in semi-natural sub-
739 alpine grasslands of Budalen, central Norway. *Biological Conservation* 87, 369-379.
- 740 Autio, J., Hicks, S., 2004. Annual variation in pollen deposition and meteorological conditions on the fell
741 Aakenustunturi in northern Finland: Potential for using fossil pollen data as a climate proxy. *Grana* 43,
742 31-47.

- 743 Behre, K.-E., 1981. The interpretation of anthropogenic indicators in pollen diagrams. *Pollen et Spores* 23, 225-
744 245.
- 745 Berglund, B.E., 1991. The Viking Age landscape. In Berglund, B.E. (Ed.), *The cultural landscape during 6000*
746 *years in southern Sweden – the Ystad project. Ecological Bulletins* 41, 82-86.
- 747 Beug, H.-J., 2004. *Leitfaden der Pollenbestimmung für Mitteleuropa und angrenzende Gebiete*. Verlag Dr.
748 Friedrich Pfeil, München.
- 749 Birks H.J.B., Birks H.H., 1980. *Quaternary Palaeoecology*. Arnold, London.
- 750 Bunting, M.J., 2003. Pollen – vegetation relationships in non-arboreal moorland taxa. *Review of Palaeobotany*
751 *and Palynology* 125, 285–298.
- 752 Bunting, M.J., Hjelle, K.L., 2010. Effect of vegetation data collection strategies on estimates of relevant source
753 area of pollen (RSAP) and relative pollen productivity estimates (relative PPE) for non-arboreal taxa.
754 *Vegetation History and Archaeobotany* 19, 365-374.
- 755 Bunting, M.J., Schofield, J.E., Edwards, K.J., 2013. Estimates of relative pollen productivity (RPP) for selected
756 taxa from southern Greenland: A pragmatic solution. *Review of Palaeobotany and Palynology* 190, 66-
757 74.
- 758 Crawley, M.J., 1997. *Life History and Environment*. In: Crawley, M.J. (Ed.), *Plant Ecology*. Blackwell Science
759 Ltd., Oxford, pp. 73-131.
- 760 de Klerk, P., Theuerkauf, M. & Joosten, H., 2017. Vegetation, recent pollen deposition, and distribution of some
761 non-pollen palynomorphs in a degrading ice-wedge polygon mire complex near Pokhodsk (NE Siberia),
762 including size-frequency analyses of pollen attributable to *Betula*. *Review of Palaeobotany and*
763 *Palynology* 238, 122-143.
- 764 Davis, O.K., 1984. Pollen frequencies reflect vegetation patterns in a great basin (U.S.A.) mountain range.
765 *Review of Palaeobotany and Palynology* 40, 295-315.
- 766 Dierssen, K., 1982. *Die wichtigsten Pflanzengesellschaften der Moore NW-Europas*. Conservatoire jardin
767 botaniques, Genève.
- 768 Ejarque, A., Miras, Y., Riera, S., 2011. Pollen and non-pollen palynomorph indicators of vegetation and
769 highland grazing activities obtained from modern surface and dung datasets in the eastern Pyrenees.
770 *Review of Palaeobotany and Palynology* 167, 123–139.
- 771 Ellenberg, H., 1988. *Vegetation Ecology of Central Europe*. Cambridge University Press, Cambridge.

- 772 Follestad, B.A., 1995. Møre og Romsdal fylke – Kwartærgeologisk kart. M 1:250.000. The Geological Survey of
773 Norway, Trondheim.
- 774 Frisvoll, A.A., Elvebakk, A., Flatberg, K.I., Økland, R.H., 1995. Sjekklister over norske mosar: vitenskapelig og
775 norsk namneverk. NINA Temahefte 4, Trondheim.
- 776 Fægri, K., Iversen, J., 1989. Textbook of Pollen Analysis. 4th edition by K. Fægri, P.E. Kaland, and K.
777 Kryzwinski. John Wiley & Sons, Chichester.
- 778 Gaillard, M.-J., Birks, H.J.B., Emanuelsson, U., Karlsson, S., Lagerås, P., Olausson, D., 1994. Application of
779 modern pollen/land-use relationships to the interpretation of pollen diagrams – reconstructions of land-
780 use history in south Sweden, 3000-0 BP. *Review of Palaeobotany and Palynology* 82, 47-73.
- 781 Grime, J.P., 2001. Plant strategies, vegetation processes, and ecosystem properties. John Wiley & Sons Ltd,
782 Chichester.
- 783 Grimm, E.C., 1990. TILIA and TILIA.GRAPH: PC spreadsheet and graphics software for pollen data. INQUA,
784 Working Group on Data-Handling Methods, Newsletter 4, 5-7.
- 785 Gunnarsdóttir, H., 1999. Postglasial vegetasjonshistorie i Nord-Gudbrandsdalen, sentrale Sør-Norge. AmS-
786 Rapport 12A, 113–144. Stavanger.
- 787 Hättestrand, M., Jensen, C., Hallsdóttir, M., Vorren, K.-D., 2008. Modern pollen accumulation rates at the north-
788 western fringe of the European boreal forest. *Review of Palaeobotany and Palynology* 151, 90-109.
- 789 Herben, T., Münzbergová, Z., Mildén, M., Ehrlén, J., Cousins, S.A.O., Eriksson, O., 2006. Long-Term Spatial
790 Dynamics of *Succisa pratensis* in a Changing Rural Landscape: Linking Dynamical Modelling with
791 Historical Maps. *Journal of Ecology* 94, 131-143.
- 792 Hicks, S., 1977. Modern pollen rain in Finnish Lapland investigated by analysis of surface moss samples. *New
793 Phytologist* 78, 715-734.
- 794 Hicks, S., 2001. The use of annual arboreal pollen deposition values for delimiting tree-lines in the landscape
795 and exploring models of pollen dispersal. *Review of Palaeobotany and Palynology* 117, 1-29.
- 796 Hjelle, K.L., 1997. Relationships between pollen and plants in human-influenced vegetation types using
797 presence-absence data in western Norway. *Review of Palaeobotany and Palynology* 99, 1-16.
- 798 Hjelle, K.L., 1998. Herb pollen representation in surface moss samples from mown meadows and pastures in
799 western Norway. *Vegetation History and Archaeobotany* 7, 79-96.
- 800 Hjelle, K.L., 1999a. Modern pollen assemblages from mown and grazed vegetation types in western Norway.
801 *Review of Palaeobotany and Palynology* 107, 55–81.

802 Hjelle, K.L. 1999b. Use of modern pollen samples and estimated pollen representation factors as aids in the
803 interpretation of cultural activity in local pollen diagrams. *Norwegian Archaeological Review* 32, 19–
804 39.

805 Hjelle, K.L., Kaland, S., Kvamme, M., Lødøen, T.K., Natlandsmyr, B., 2012. Ecology and long-term land-use,
806 palaeoecology and archaeology – the usefulness of interdisciplinary studies for knowledge-based
807 conservation and management of cultural landscapes. *International Journal of Biodiversity Science,*
808 *Ecosystem Services & Management* 8, 321-337.

809 Hulme, P.D., Pakeman, R.J., Torvell, L., Fisher, J.M. and Gordon, I.J., 1999. The effects of controlled sheep
810 grazing on the dynamics of upland *Agrostis–Festuca* grassland. *Journal of Applied Ecology* 36, 886-
811 900.

812 Janssen, C.R., 1966. Recent pollen spectra from the deciduous and coniferous-deciduous forests of northeastern
813 Minnesota: A study in pollen dispersal. *Ecology* 47, 804-825.

814 Jiménez-Alfaro, B., Hájek, M., Ejrnaes, R., Rodwell, J., Pawlikowski, P., Weeda E.J., Laitinen, J., Moen, A.,
815 Bergamini, A., Aunina, L., Sekulová, L., Tahvanainen, T., Gillet, F., Jandt, U., Dítě, D., Hájková, P.,
816 Corriol, G., Kondelin, H., Díaz, T.E., 2014. Biogeographic patterns of base-rich fen vegetation across
817 Europe. *Applied Vegetation Science* 17, 367–380.

818 Klanderud, K., Totland, Ø., 2005. The relative importance of neighbours and abiotic environmental conditions
819 for population dynamic parameters of two alpine plant species. *Journal of Ecology* 93, 493-501.

820 Klimešová, J., Latzel, V., de Bello, F., van Groenendael, J.M., 2008. Plant functional traits in studies of
821 vegetation changes in response to grazing and mowing: towards a use of more specific traits. *Preslia* 80,
822 245-253.

823 Lid, J., Lid, D.T., 2005. *Norsk flora*. 7th edition Elven, R. (Ed.). Det Norske Samlaget, Oslo.

824 Lyngstad, A., Moen, A., Pedersen, B., 2016. Flowering in the rich fen species *Eriophorum latifolium* depends on
825 climate and reproduction in the previous year. *Wetlands* 37, 1–13.

826 Mazier, F., Galop, D., Brun, C., Buttler, A., 2006. Modern pollen assemblages from grazed vegetation in the
827 western Pyrenees, France: a numerical tool for more precise reconstruction of past cultural landscapes.
828 *The Holocene* 16, 91-103.

829 Moe, D., 1974. Identification key for trilete microspores of Fennoscandian Pteridophyta. *Grana* 14, 132-142.

830 Moen, A., 1990. The plant cover of the boreal uplands of Central Norway. I. Vegetation ecology of Sølendet
831 nature reserve; haymaking fens and birch woodlands. *Gunneria* 63, 1-451.

- 832 Moen, A., 1995. Vegetational changes in boreal rich fens induced by haymaking; management plan for the
833 Sølendet Nature Reserve. In: Wheeler, B.D., Shaw, S.C., Fojt, W.J., Robertson, R.A. (Eds.), Restoration
834 of temperate wetlands. John Wiley & Sons, Chichester, pp. 167-181.
- 835 Moen, A., 1999. National Atlas of Norway: Vegetation. Norwegian Mapping Authority, Hønefoss.
- 836 Moen, A., 2000. Botanisk kartlegging og plan for skjøtsel av Tågdalen naturreservat i Surnadal. NTNU
837 Vitenskapsmuseet, Rapport botanisk serie 2000-7, 1-45.
- 838 Moen, A., Nilsen, L.S., Øien, D.-I., Arnesen, T., 1999. Outlying haymaking lands at Sølendet, central Norway:
839 effects of scything and grazing. *Norsk Geografisk Tidsskrift* 53, 93-102.
- 840 Moen, A., Lyngstad, A., Øien, D.-I., 2012. Boreal rich fen vegetation formerly used for haymaking. *Nordic*
841 *Journal of Botany* 30, 226–240.
- 842 Moen, A., Lyngstad, A., Øien, D.-I., 2015. Hay crop of boreal rich fen communities traditionally used for
843 haymaking. *Folia Geobotanica* 50, 25-38.
- 844 Moore, P.D., Webb, J.A., Collinson, M.E., 1991. *Pollen Analysis*. Blackwell Scientific Publications, Oxford.
- 845 Mueller-Dombois, D., Ellenberg, H., 1974. *Aims and Methods of Vegetation Ecology*. John Wiley and Sons,
846 New York.
- 847 Mulder, C., Janssen, C.R., 1999. Occurrence of pollen and spores in relation to present-day vegetation in a Dutch
848 heathland area. *Journal of Vegetation Science* 10, 87–100.
- 849 Nielsen, A.B., Møller, P.F., Giesecke, T., Stavngaard, B., Fontana, S.L., Bradshaw, R.H.W., 2010. The effect of
850 climate conditions on inter-annual flowering variability monitored by pollen traps below the canopy in
851 Draved forest, Denmark. *Vegetation History and Archaeobotany* 19, 309–323.
- 852 Øien, D.-I., Moen, A., 2001. Nutrient limitation in boreal plant communities and species influenced by scything.
853 *Applied Vegetation Science* 4, 197-206.
- 854 Pardoe, H.S., 1996. Micro-scale patterns of modern pollen deposition within three alpine plant communities.
855 *New Phytologist* 132, 327-341.
- 856 Pardoe, H.S., Giesecke, T., van der Knaap, W.O., Svitavská-Svobodová, H., Kvavadze, E.V., Panajiotidis, S.,
857 Gerasimidis, A., Pidek, I.A., Zimny, M., Święta-Musznicka, J., Latałowa, M., Noryśkiewicz, M.A.,
858 Bozilova, E., Tonkov, S., Filipova-Marinova, M.V., van Leeuwen, J.F.N., Kalniņa, L., 2010.
859 Comparing pollen spectra from modified Tauber traps and moss samples: examples from a selection of
860 woodlands across Europe. *Vegetation History and Archaeobotany* 19, 271-283.

861 R Development Core Team, 2010. R: A language and environment for statistical computing. R Foundation for
862 Statistical Computing. Vienna, Austria.

863 Randall, R.E., Andrew, R., West, R.G., 1986. Pollen catchment in relation to local vegetation: Ceann Ear,
864 Monach Isles N.N.R., Outer Hebrides. *New Phytologist* 104, 271-310.

865 Räsänen, S., Hicks, S., Odgaard, B.V., 2004. Pollen deposition in moss and in a modified «Tauber trap» from
866 Hailuoto, Finland: what exactly do the mosses record? *Review of Palaeobotany and Palynology* 129,
867 103-116.

868 Rowell, T.A., Guarino, L., Harvey, H.J., 1985. The experimental management of vegetation at Wicken Fen,
869 Cambridgeshire. *Journal of Applied Ecology* 22, 217-227.

870 Rybníček, K., 1985. A Central-European approach to the classification of mire vegetation. *Aquilo Seria Botanica*
871 21, 19–31.

872 Sigmond, E.M.O., Gustavson, M., Roberts, D., 1984. *Berggrunnskart over Norge*. The Geological Survey of
873 Norway, Trondheim.

874 Sjörs, H., 1948. Myrvegetation i Bergslagen. *Acta Phytogeographica Suecica* 21, 1-299.

875 Sletvold, N., Øien, D.-I., Moen, A., 2010. Long-term influence of mowing on population dynamics in the rare
876 orchid *Dactylorhiza lapponica*: The importance of recruitment and seed production. *Biological*
877 *Conservation* 143, 747-755.

878 Solem, T.L., Aune, E.I., Daverdin, M., Hassel, K., Sjögren, P.J.E., Stenvik, L., Tretvik, A.M., Øien, D.-I.,
879 Austrheim, G., 2012. Long-term land-use and landscape dynamics in Budalen, central Norway.
880 *International Journal of Biodiversity Science, Ecosystem Services & Management*, 8, 351-359.

881 Sugita, S., 1994. Pollen representation of vegetation in Quaternary sediments: theory and method in patchy
882 vegetation. *Journal of Ecology* 82, 881-897.

883 ter Braak, C.J.F., Šmilauer, P., 2002. *CANOCO reference manual and CanoDraw for Windows user's guide:*
884 *software for canonical community ordination (version 4.5)*. New York, USA.

885 Tretvik, A.M., Krogstad, K., 1999. Historisk studie av utmarkas betydning økonomisk og sosialt innen
886 Tågdalens naturreservat for Dalsegg-grenda i Øvre Surnadal. NTNU Vitenskapsmuseet, Rapport
887 botanisk serie, 1999-2, 1-38.

888 Waller, M., Binney, H.A., Bunting, M.J., Armitage, R. 2005. The interpretation of fen carr pollen diagrams:
889 pollen-vegetation relationships within fen carr. *Review of Palaeobotany and Palynology* 133, 179-202.

- 890 Waller, M., Carvalho, F., Grant, M.J., Bunting, M.J., Brown, K. 2017. Disentangling the pollen signal from fen
891 systems: Modern and Holocene studies from southern and eastern England. *Review of Palaeobotany*
892 *and Palynology* 238, 15–33.
- 893 Webb, T., III, Lasleski, R.A., Bernabo, J.C., 1978. Sensing vegetational patterns with pollen data: choosing the
894 data. *Ecology* 59, 1151-1163.
- 895

896 Figure legends

897

898 Fig. 1. The location of the two study areas, Tågdalen in the inner fjord area and the inland
899 area Sølendet. Vegetation sections in central Norway after Moen (1999).

900

901 Fig. 2. The location of investigated localities in a) Tågdalen Nature Reserve and b) Sølendet
902 Nature Reserve. The reserves are marked with black solid borders. Numbering of the plots
903 follows the permanent plots in long-term studies from the reserves (Moen, 1990, 2000).

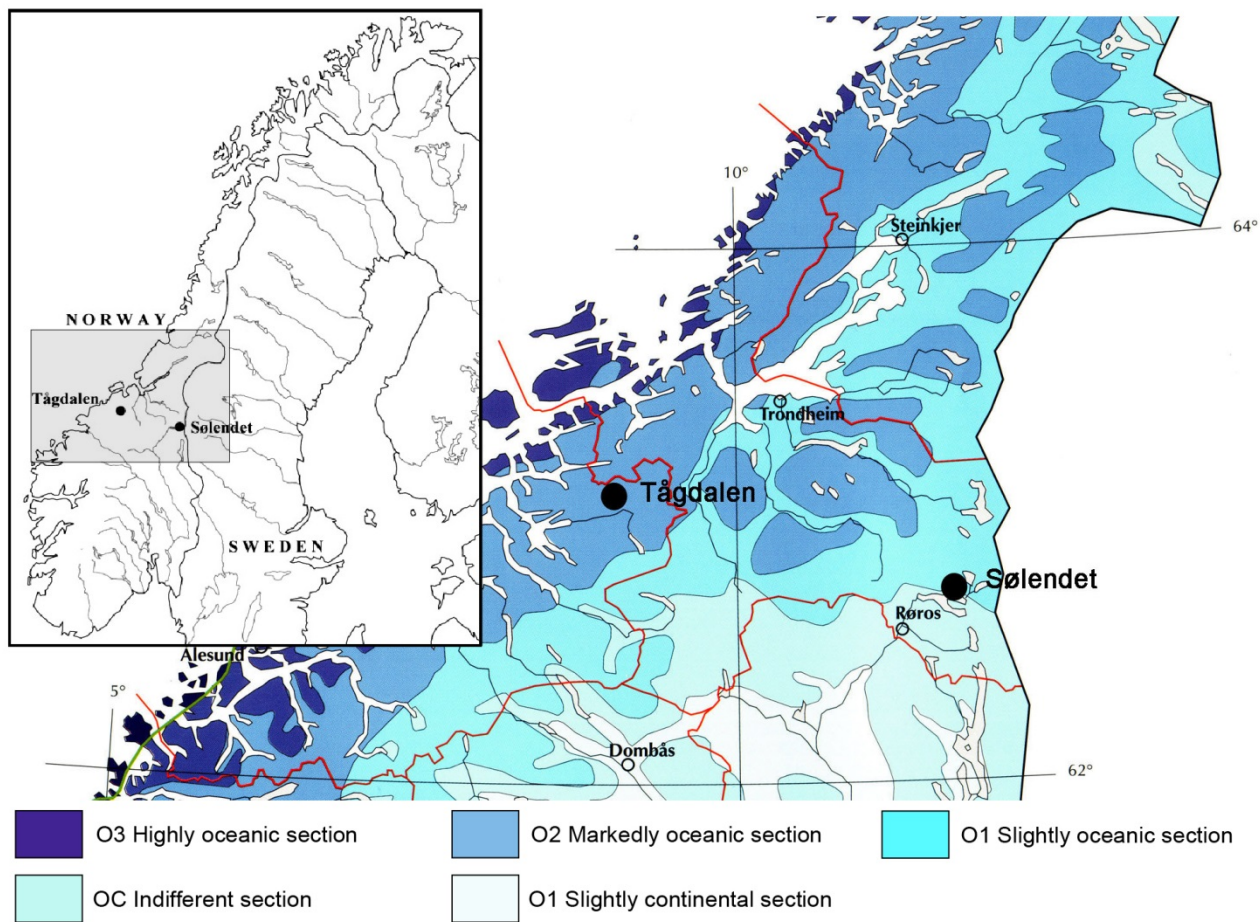
904

905 Fig. 3. Schematic presentation of a typical locality with permanent plots, investigated plots of
906 1 m² for the present study and collection points for moss samples. A= mown biennially, C=
907 mown quadrennially and B= unmown.

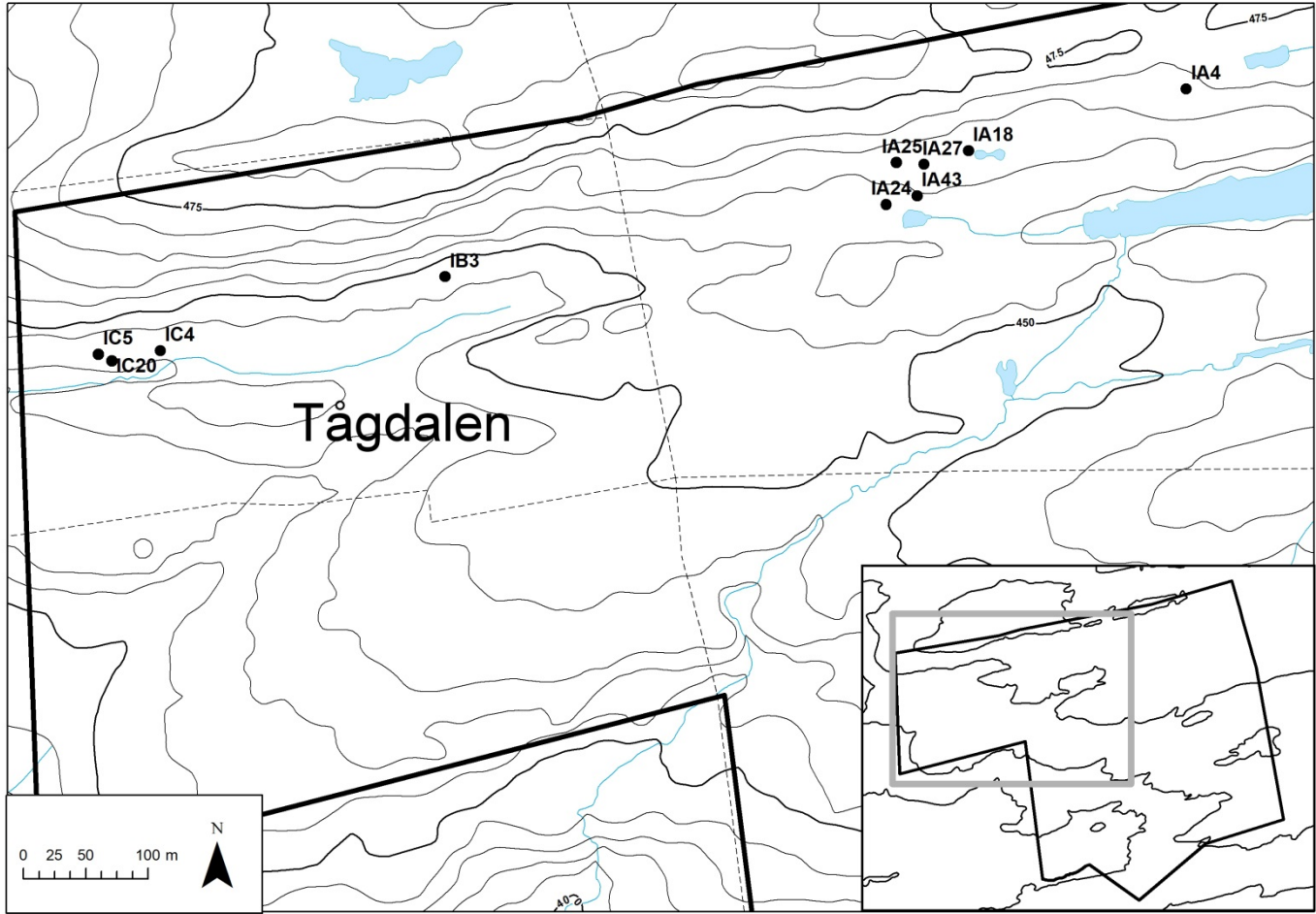
908

909 Fig. 4. PCA showing a) vegetation samples, pollen samples and environmental variables.
910 Pollen samples and environmental variables are treated as passive samples and variables
911 respectively. Veg. = vegetation samples, Pollen = pollen samples, 2nd = mown biennially, 4th
912 = mown quadrennially. Grey triangles = environmental variables, three mowing regimes, b)
913 plant species grouped into pollen taxa according to Table 2, based on 26 plots from Tågdalen
914 and 23 plots from Sølendet. Pollen types close to the center are not shown to increase
915 readability. Abbreviations: t = type, *Menyanthes tri* = *Menyanthes trifoliata*, *Saxifraga opp* =
916 *Saxifraga oppositifolia*.

917



918
 919 Fig. 1.
 920



921
922
923
924

Fig. 2a.

925 **Table 1**
 926 Characteristics of the two study areas. Climatic data from Norwegian Meteorological Institute
 927 (see Lyngstad et al., 2016, including Supplementary Material) for the study areas at Tågdalen
 928 nature reserve and Sølendet nature reserve. Vegetation zones and sections after Moen (1999).
 929

	Tågdalen	Sølendet
Latitude (N)	63°03'	62°40'
Longitude (E)	9°05'	11°50'
Altitude (m a.s.l.)	440-460	715-785
Mean annual precipitation (mm)	1583	637
Mean annual temperature (°C)	3.0	0.6
Mean July temperature (°C)	11.2	10.5
Mean January temperature (°C)	- 2.7	- 9.5
Vegetation zone	middle/ northern boreal	middle/ northern boreal
Vegetation section	markedly oceanic (O2)	indifferent (OC)/slightly oceanic (O1)

930
 931

932 **Table 2**
 933 The 26 pollen and spore taxa with their corresponding 57 plant species present in the studied
 934 localities. Only taxa registered in both vegetation and pollen samples are included. T = pollen
 935 or plant only present in Tågdalen; S = pollen or plant only present in Sølendet. Nomenclature
 936 for pollen taxa follows Fægri and Iversen (1989), plant taxa follow Frisvoll et al. (1995) and
 937 Lid and Lid (2005).

Pollen/spore taxa	Plant species in the vegetation plots
<i>Anemone</i>	<i>Anemone nemorosa</i> ^T
Asteraceae Cichorioideae	<i>Crepis paludosa</i> ^S , <i>Leontodon autumnalis</i> ^S
<i>Betula</i>	<i>Betula nana</i> , <i>Betula nana</i> x <i>pubescens</i> , <i>Betula pubescens</i>
<i>Bistorta vivipara</i> -type	<i>Bistorta vivipara</i>
<i>Calluna vulgaris</i>	<i>Calluna vulgaris</i> ^S
Cyperaceae	<i>Carex</i> spp., <i>Eleocharis quinqueflora</i> , <i>Eriophorum</i> spp., <i>Kobresia simpliciuscula</i> ^S , <i>Schoenus ferrugineus</i> ^T , <i>Trichophorum</i> spp.,
<i>Drosera rotundifolia</i> -type ^T	<i>Drosera longifolia</i> ^T , <i>Drosera rotundifolia</i> ,
<i>Dulichium</i> -type	<i>Carex capillaris</i> , <i>Carex chordorrhiza</i> ^T , <i>Carex dioica</i> , <i>Carex echinata</i> , <i>Carex flava</i> , <i>Carex flava</i> x <i>hostiana</i> ^S , <i>Carex hostiana</i> , <i>Carex lasiocarpa</i> , <i>Carex limosa</i> ^T , <i>Carex nigra</i> , <i>Carex panicea</i> , <i>Carex pauciflora</i> ^T , <i>Carex rostrata</i> , <i>Carex vaginata</i> ^S
Ericaceae	<i>Andromeda polifolia</i> , <i>Calluna vulgaris</i> ^S , <i>Erica tetralix</i> ^T , <i>Vaccinium uliginosum</i> ssp. <i>uliginosum</i>
<i>Eriophorum</i> -type	<i>Eriophorum angustifolium</i> , <i>Eriophorum latifolium</i> , <i>Eriophorum vaginatum</i> ^S , <i>Trichophorum alpinum</i> , <i>Trichophorum cespitosum</i> ssp. <i>cespitosum</i>
<i>Galium</i> -type ^T	<i>Galium boreale</i>
<i>Juniperus communis</i>	<i>Juniperus communis</i>
<i>Menyanthes trifoliata</i> ^T	<i>Menyanthes trifoliata</i> ^T
<i>Pedicularis</i> -type	<i>Pedicularis oederi</i> ^S , <i>Pedicularis palustris</i>
<i>Pinguicula</i> -type ^S	<i>Pinguicula vulgaris</i> ^S
Poaceae	<i>Deschampsia cespitosa</i> ^S , <i>Festuca ovina</i> ^S , <i>Festuca vivipara</i> ^T , <i>Molinia caerulea</i> , <i>Nardus stricta</i>
<i>Potentilla</i> -type	<i>Potentilla erecta</i>
<i>Rhinanthus</i> -type ^S	<i>Bartsia alpina</i> ^S , <i>Euphrasia wettsteinii</i>
Rosaceae	<i>Potentilla erecta</i>
<i>Salix</i> ^S	<i>Salix glauca</i> ^S
<i>Saxifraga oppositifolia</i> -type	<i>Saxifraga aizoides</i> ^S
<i>Selaginella selaginoides</i>	<i>Selaginella selaginoides</i>
<i>Solidago</i> -type	<i>Solidago virgaurea</i> ^S
<i>Sphagnum</i>	<i>Sphagnum warnstorffii</i> , <i>Sphagnum</i> Section <i>Subsecunda</i>
<i>Succisa pratensis</i>	<i>Succisa pratensis</i>
<i>Thalictrum</i>	<i>Thalictrum alpinum</i>

938
 939

940 **Table 3**
 941 Number of plots with plant recorded (n), mean plant cover in vegetation plots based on 26
 942 plots from Tågdalen (T) and 23 from Sølendet (S), and differences in cover between plots
 943 mown biennially (M2) and unmown (U). ^T = only from Tågdalen; ^S = only from Sølendet; – =
 944 not present. The difference in % values displays absolute differences on a % scale between
 945 plots mown biennially and unmown plots.

Plant taxa	n	T, mean %	S, mean %	Difference in % values M2 – U (T & S)
<i>Bartsia alpina</i>	29	2.5	0.5	-1.6
<i>Betula nana</i>	16	0.01	2.9	-5.7
<i>Bistorta vivipara</i>	19	0.1	0.6	-0.5
<i>Carex capillaris</i>	20	0.7	0.8	0.3
<i>Carex dioica</i>	40	2.2	1.1	2.2
<i>Carex echinata</i>	9	0.7	0.01	2.1
<i>Carex flava</i>	35	2.6	6.1	5.9
<i>Carex hostiana</i>	22	4.3	0.01	4.9
<i>Carex limosa</i> ^T	25	2.8	–	2.4
<i>Carex nigra</i>	11	0.02	1.8	2.1
<i>Carex panicea</i>	48	2.2	4.0	1.2
<i>Carex rostrata</i>	30	3.1	2.3	4.8
<i>Drosera longifolia</i> ^T	13	2.0	–	4.0
<i>Eriophorum angustifolium</i>	43	0.7	0.8	0.9
<i>Eriophorum latifolium</i>	44	7.8	6.4	5.7
<i>Eriophorum vaginatum</i> ^S	10	–	0.3	-0.5
<i>Euphrasia wettsteinii</i>	25	0.8	1.4	0.0
<i>Festuca ovina</i> ^S	12	–	0.7	-0.7
<i>Menyanthes trifoliata</i>	9	0.4	1.6	-5.4
<i>Molinia caerulea</i>	49	8.3	7.4	-13.6
<i>Pedicularis oederi</i> ^S	11	–	0.5	-0.3
<i>Pedicularis palustris</i>	26	0.2	3.0	3.1
<i>Pinguicula vulgaris</i>	13	1.7	0.9	2.6
<i>Potentilla erecta</i>	44	9.2	5.2	0.4
<i>Selaginella selaginoides</i>	43	1.5	1.9	0.2
<i>Succisa pratensis</i>	32	3.5	1.6	-1.9
<i>Thalictrum alpinum</i>	49	10.8	13.9	10.6
<i>Trichophorum alpinum</i>	22	1.7	0.1	-0.4
<i>Trichophorum cespitosum</i> <i>ssp. cespitosum</i>	49	18.4	14.5	-1.9

946
 947

948 **Table 4**
949 Results (p-values) from paired comparisons of differences in a) mean percentage cover of
950 plant taxa and b) sum of mean percentage cover of pollen taxa between management regimes
951 in Tågdalen and Sølendet using the Wilcoxon signed rank test. The test was run for plots
952 mown biennially and unmown plots (M2 & U), plots mown biennially and quadrennially (M2
953 & M4), and plots mown quadrennially and unmown plots (M4 & U). Numbers in bold display
954 values statistically significant at a 0.05-level. -- = not enough plots to compute the test. The
955 rightmost column displays indices of difference (ID), range +1.0 to -1.0, between plots mown
956 biennially and unmown (M2 & U). Positive values refer to taxa with an increase in plant
957 cover with mowing, negative values vice versa.

a) Plant taxa	p-values			ID
	M2 & U	M2 & M4	M4 & U	M2 & U
<i>Molinia caerulea</i>	<0.0001	0.047	0.010	-0.95
<i>Carex dioica</i>	<0.0001	0.25	0.22	0.89
<i>Thalictrum alpinum</i>	<0.0001	0.50	0.06	0.90
<i>Pinguicula vulgaris</i>	0.001	0.031	0.38	0.65
<i>Carex hostiana</i>	0.002	0.50	0.06	1.00
<i>Carex flava</i>	0.002	>0.99	0.13	0.60
<i>Eriophorum latifolium</i>	0.003	0.50	0.06	0.72
<i>Eriophorum angustifolium</i>	0.012	0.06	0.13	0.44
<i>Pedicularis palustris</i>	0.017	>0.99	0.50	0.58
<i>Betula nana</i>	0.027	--	--	-0.60
<i>Succisa pratensis</i>	0.036	>0.99	>0.99	-0.38
<i>Carex limosa</i> ^T	0.047	>0.99	0.38	0.60
<i>Carex echinata</i>	0.06	--	--	0.83
<i>Drosera longifolia</i> ^T	0.06	>0.99	0.25	0.83
<i>Bartsia alpina</i>	0.18	0.63	0.25	-0.40
<i>Carex nigra</i>	0.19	--	--	0.50
<i>Eriophorum vaginatum</i> ^S	0.22	--	--	-0.57
<i>Carex panicea</i>	0.23	0.25	0.25	0.40
<i>Bistorta vivipara</i>	0.27	--	0.38	-0.40
<i>Carex rostrata</i>	0.31	0.06	>0.99	0.25
<i>Trichophorum alpinum</i>	0.34	0.31	0.63	-0.29
<i>Festuca ovina</i> ^S	0.38	--	--	-0.29
<i>Trichophorum cespitosum</i> <i>ssp. cespitosum</i>	0.44	0.38	0.50	-0.15
<i>Menyanthes trifoliata</i>	0.50	0.75	0.88	-0.25
<i>Carex capillaris</i>	0.67	--	--	0.10
<i>Euphrasia wettsteinii</i>	0.70	--	--	-0.07
<i>Selaginella selaginoides</i>	0.72	0.06	0.25	0.05
<i>Potentilla erecta</i>	0.87	0.13	0.31	0.11
<i>Pedicularis oederi</i> ^S	0.94	--	--	-0.17
b) Plant taxa in pollen				
taxa group				
<i>Dulichium-type</i>	<0.0001	0.047	0.039	1.00
Cyperaceae	<0.0001	0.06	0.012	1.00
Poaceae	<0.0001	0.09	0.008	-0.95
<i>Eriophorum</i> spp.	0.002	0.14	0.039	0.60
<i>Betula</i>	0.002	--	0.38	-0.69
<i>Pedicularis-type</i>	0.023	>0.99	0.13	0.38
<i>Drosera</i> . ^T	0.031	0.38	0.25	0.75

958	<i>Eriophorum</i> -type (incl. <i>Trichophorum</i>)	0.040	0.77	0.35	0.30
-----	---	--------------	------	------	------

959 **Table 5**
 960 Mean pollen percentages of local terrestrial pollen and spores from the two study areas,
 961 Tågdalen and Sølendet, for each land-use regime. SE= standard error; n= number of samples.
 962 * displays pollen taxa included in the local terrestrial pollen sum. Cyperaceae includes
 963 *Dulichium*-type and *Eriophorum*-type, Ericaceae includes *Calluna vulgaris*, Rosaceae
 964 includes *Potentilla*-type. *Menyanthes trifoliata*, Ericaceae, *Drosera rotundifolia*-type,
 965 *Saxifraga oppositifolia*-type and *Galium*-type from Tågdalen and *Anemone*, *Saxifraga*
 966 *oppositifolia*-type, *Bistorta vivipara*-type, *Pinguicula*-type and *Rhinanthus*-type from
 967 Sølendet, were found in few plots and with low percentages and are not included.
 968

Pollen/spore taxa	Mown biennially			Mown quadrennially			Unmown		
	n (of 10)	Mean %	SE	n (of 6)	Mean %	SE	n (of 10)	Mean %	SE
Tågdalen									
Cyperaceae	10	80.0	1.8	6	78.2	3.2	10	80.8	1.4
<i>Betula</i>	10	32.7	1.9	6	29.7	3.5	10	29.1	1.3
<i>Dulichium</i> -type*	10	25.5	1.3	6	25.2	1.6	10	26.7	2.4
<i>Eriophorum</i> -type*	10	16.2	0.6	6	16.1	2.3	10	9.0	1.3
<i>Selaginella selaginoides</i>	10	15.1	1.5	6	11.0	2.2	10	17.4	3.7
Poaceae*	10	7.8	0.5	6	9.3	1.2	10	9.4	0.7
<i>Thalictrum</i> *	10	4.8	1.0	6	5.9	1.6	10	3.2	0.6
Rosaceae	9	2.4	0.5	6	2.6	0.4	10	2.9	0.3
<i>Potentilla</i> -type*	8	1.4	0.3	5	2.0	0.3	10	2.5	0.3
<i>Juniperus communis</i>	9	0.9	0.2	4	0.5	0.2	9	0.7	0.1
<i>Sphagnum</i>	7	0.6	0.2	5	0.6	0.1	7	0.6	0.2
<i>Solidago</i> -type*	3	0.4	0.2	1	0.1	0.1	3	0.1	0.05
<i>Succisa pratensis</i> *	5	0.2	0.07	1	0.2	0.1	3	0.2	0.1
<i>Anemone</i> *	4	0.2	0.08	1	0.06	0.04	4	0.2	0.07
<i>Pedicularis</i> -type*	3	0.2	0.09	3	0.3	0.1	0	0	0
<i>Menyanthes trifoliata</i> ^T *	1	0.1	0.03	1	0.08	0.05	2	0.05	0.03
Sølendet									
Cyperaceae	10	78.5	4.1	3	90.4	2.2	10	83.7	2.8
<i>Betula</i>	10	32.4	3.0	3	18.2	2.6	10	30.2	2.9
<i>Eriophorum</i> -type*	10	28.9	2.2	3	35.9	1.9	10	14.9	1.1
<i>Selaginella selaginoides</i>	10	19.2	3.1	3	9.6	1.2	10	11.3	1.8
<i>Dulichium</i> -type *	10	18.4	1.2	3	21.6	2.4	10	24.4	1.2
<i>Thalictrum</i> *	10	12.6	3.9	3	5.4	2.6	10	3.9	1.1
Poaceae*	10	4.6	0.6	3	2.7	0.5	10	7.2	1.3
<i>Juniperus communis</i>	8	1.0	0.2	3	0.5	0.1	8	0.6	0.1
<i>Pedicularis</i> -type*	7	0.6	0.1	2	0.6	0.3	7	0.4	0.1
Asteraceae CichorioideaeS*	9	0.5	0.1	1	0.1	0.1	5	0.5	0.2
<i>Sphagnum</i>	6	0.4	0.1	3	0.7	0.2	10	0.5	0.1
Ericaceae	5	0.3	0.1	1	0.2	0.2	6	0.3	0.1
<i>Calluna vulgaris</i>	5	0.3	0.1	1	0.2	0.2	4	0.2	0.1
<i>Salix</i> ^S	6	0.3	0.1	2	0.3	0.1	7	0.4	0.1
Rosaceae	8	1.0	0.2	0	0	0	9	2.1	0.5
<i>Potentilla</i> -type*	7	0.7	0.2	0	0	0	9	2.0	0.5
<i>Solidago</i> -type*	4	0.2	0.1	0	0	0	4	0.3	0.1
<i>Succisa pratensis</i> *	4	0.2	0.1	0	0	0	2	0.6	0.3

969

970 **Table 6**
 971 Correlation between plant cover (transformed to pollen taxa) percentages and pollen
 972 percentages, and indices of association (A, U, O) based on presence/absence data.
 973 Abbreviations: T = test only from Tågdalen; S = only from Sølendet; ρ (rho) = Spearman rank-
 974 correlation coefficient. Correlation coefficient significant at $p < 0.05$ in bold. The indices of
 975 association were classified into SAT = strongly associated types, AT = associated types,
 976 WAT = weakly associated types, ORT = over-represented type, URT = under-represented
 977 type and UT = unassociated type. See section 3.4.5. for further description. *Anemone* (WAT),
 978 *Calluna vulgaris* (ORT), *Juniperus communis* (WAT), *Salix* (UT) and *Solidago*-type (WAT)
 979 were present in too few plots with plant or pollen to conduct the test.

Pollen taxon	ρ	A	U	O	Association
Rosaceae	0.47	0.83	0.11	0.07	SAT
<i>Potentilla</i> -type	0.44	0.8	0.16	0.05	SAT
<i>Pedicularis</i> -type	0.43	0.59	0.34	0.14	AT
<i>Succisa pratensis</i>	0.43	0.38	0.59	0.13	WAT
<i>Thalictrum</i>	0.35	1	0	0	SAT
<i>Menyanthes trifoliata</i>	0.31	0.25	0.67	0.5	WAT
<i>Selaginella selaginoides</i>	0.30	0.88	0	0.12	SAT
<i>Eriophorum/Eriophorum spp.</i>	0.25	0.98	0	0.02	SAT
Poaceae/ <i>Molinia caerulea</i>	0.25	1	0	0	SAT
Cyperaceae	0.24	1	0	0	SAT
<i>Drosera rotundifolia</i> -type ^T	0.21	0.06	0.94	0	URT
Poaceae	0.20	1	0	0	SAT
Asteraceae Cichorioideae ^S	0.18	0.19	0.25	0.8	WAT
<i>Bistorta vivipara</i> -type ^T	0.11	0.06	0.94	0	URT
<i>Betula</i>	0.05	0.43	0	0.57	ORT
<i>Rhinanthus</i> -type ^S	0.04	0.13	0.87	0	URT
Ericaceae	0.02	0.19	0.74	0.57	WAT
<i>Galium</i> -type ^T	-0.05	0	1	1	UT
<i>Sphagnum</i>	-0.13	0.21	0.11	0.79	WAT
<i>Dulichium</i> -type	-0.16	1	0	0	SAT
<i>Pinguicula</i> -type ^S	-0.22	0	1	1	UT
<i>Saxifraga oppositifolia</i> -type	-0.96	0.13	0.88	0	URT
<i>Eriophorum</i> -type	-0.01	1	0	0	SAT

980

981 **Table 7**
 982 Pollen indicator taxa for management regimes in boreal rich fens based on the results from the
 983 Wilcoxon signed rank test (relationship between vegetation and management regime) and the
 984 Spearman rank-correlation coefficient (relationship between plant and pollen percentages).
 985 The effect of traditional mowing on the relative vegetation cover (plants expressed as pollen
 986 taxa) in fens is displayed as: + positive effect of traditional mowing, • no effect, - negative
 987 effect). Results from earlier studies in the study areas are displayed in brackets (after Moen,
 988 1990, 1995; Moen et al., 1999); see text for further details.
 989 **Pedicularis*-type pollen mainly increases with mowing, but single species are reduced with
 990 mowing.

Pollen-vegetation relationship	Pollen/spore taxa	Effect of mowing
Very good pollen indicator	<i>Thalictrum</i>	+
Good pollen indicators	Cyperaceae	+
	<i>Eriophorum</i> -type	+
	Poaceae	-
	<i>Succisa pratensis</i>	-
	<i>Pedicularis</i> -type	+*
Potential pollen indicators	<i>Potentilla</i> -type	(-)
	Rosaceae	(-)
	<i>Selaginella selaginoides</i>	(+)
Pollen type reflecting unidentified factors in present study	<i>Menyanthes trifoliata</i>	(+•)
Pollen types reflecting plant cover in earlier studies	<i>Anemone</i>	
	Asteraceae Cichorioideae (<i>Leontodon</i> spp./ <i>Crepis</i> spp.)	(+/-)
	<i>Betula</i>	(-)
	<i>Bistorta vivipara</i> -type	(+)
	<i>Calluna vulgaris</i>	(-)
	<i>Dulichium</i> -type	(+•/-)
	<i>Drosera rotundifolia</i> -type	(+)
	Ericaceae	(-)
	<i>Galium</i> -type	
	<i>Juniperus communis</i>	(-)
	<i>Pinguicula</i> -type	(+)
	<i>Rhinanthus</i> -type (<i>Euphrasia</i> spp./ <i>Bartsia</i> spp.)	(+)
	<i>Salix</i>	(+/-)
	<i>Saxifraga oppositifolia</i> -type	(+)
	<i>Sphagnum</i>	(-)
<i>Solidago</i> -type	(-)	

991
 992